

Patch-burn grazing (PBG) as a livestock management alternative for fire-prone ecosystems of North America

J.D. Scasta^{1*}, E.T. Thacker², T.J. Hovick³, D.M. Engle⁴, B.W. Allred⁵, S.D. Fuhlendorf⁴ and J.R. Weir⁴

¹Department of Ecosystem Science and Management, University of Wyoming, Agriculture C 2004, Laramie, Wyoming 82071, USA.

²Wildland Resources Department, Utah State University, 5230 Old Main Hill, Logan, Utah 84322, USA.

³School of Natural Resource Sciences, North Dakota State University, 202 Hultz Hall, Fargo, North Dakota 58108, USA.

⁴Department of Natural Resource Ecology and Management, Oklahoma State University, 008C Agricultural Hall, Stillwater, Oklahoma 74078, USA.

⁵College of Forestry and Conservation, University of Montana, 32 Campus Drive, Missoula, Montana 59812, USA.

*Corresponding author: jscasta@uwyo.edu

Accepted 25 September 2015; First published online 27 October 2015

Review Article

Abstract

Many rangelands of the world are fire dependent and display a strong interaction between fire and grazing on animal behavior, productivity and ecosystem processes. The application of this fire–grazing interaction as patch-burn grazing (PBG) has recently been promoted in North America to conserve biodiversity and as an alternative for livestock management in fire-prone ecosystems to enhance forage quality and other production benefits. PBG is functionally applied by burning spatially and temporally discrete patches to allow livestock to choose where and when to graze. However, considering that the primary intent of PBG in fire-dependent ecosystems has been for the conservation of biodiversity, we synthesized the peer-reviewed literature to assess PBG as an alternative strategy for livestock management in fire-prone ecosystems. We reviewed the literature to assess PBG as an alternative livestock management approach to optimize animal production and conserve biodiversity in fire-prone ecosystems. We reviewed the results of 83 studies that focused on two main areas: (1) livestock production and inputs and (2) maintaining or improving ecosystem functioning and biodiversity to support sustainable livestock production. PBG can optimize cattle production by offsetting input costs such as supplemental feed, insecticides, herbicides, mechanical brush control, veterinary costs and cross-fencing. PBG can also maintain native herbaceous plant communities that are the resource base for cattle grazing enterprises by reducing woody plant encroachment, stimulating above- and below-ground biomass of native perennial grasses, enhancing nutrient cycling and optimizing plant diversity. PBG creates a habitat mosaic critical for many trophic levels of wildlife, particularly grassland birds, which are currently in decline. Further research is needed to clarify the potential environmental gradients defining applicability of PBG, economic outcomes of PBG, potential gastro-intestinal parasite control with PBG and other metrics of animal production. Overall, PBG is a viable management approach to improve productivity and biodiversity in fire-regulated grassland ecosystems in a manner supported by both fire and grazing disturbances. This is especially true when these communities have other organisms that depend on periodic disturbance and interaction with large animal grazing and is supported by ample empirical research.

Key words: beef cattle, biodiversity, input costs, patch-burning, pyric-herbivory, sustainability

Introduction

Many of the world's naturally occurring ecosystems, such as grasslands, savannahs and shrublands are considered fire dependent¹ and are important for livestock

production. The designation as a fire-dependent ecosystem is because regularly occurring fires create a frequent disturbance that regulates ecological patterns and processes. In many fire-dependent ecosystems, this disturbance pattern included the response of large herbivores

that were attracted to the nutritious regrowth of recently burned areas after fires removed old standing plant material^{2–7}. This ecological interaction of fire and grazing, termed pyric-herbivory or fire-driven grazing, results in a shifting mosaic of landscape patterns which increases broad-scale heterogeneity⁸. The fire regime and the spatio-temporal variability of disturbance patterns maintained grassland by stimulating perennial grasses, shifting competitive interactions and preventing woody plant encroachment. For these reasons, the coupled interaction of fire and grazing has been suggested as an equally important driver of central North American grasslands as climate and soil processes⁶. This fire-grazing phenomena is not unique to North America only, as evidence suggests many vegetation types and organisms of Africa, Asia and Australia are also highly dependent on this ecological interaction as well^{2,4–7}.

Unique to North America though, is that over the last two decades, ecologists and conservation-focused organizations such as The Nature Conservancy have tried to restore the interaction of fire and grazing through the use of patch-burn grazing or PBG^{9–12} (Fig. 1). Settlement patterns by non-indigenous people had led to the suppression of fire and extirpation of bison, effectively removing these disturbances from the landscape¹³. Settlers were drawn to the vast expanses of productive forages that could be the foundation of a burgeoning livestock industry in the western USA. Conventional livestock production in these fire-dependent ecosystems has replaced bison with cattle and sought uniformity in grazing patterns and plant communities. Incidentally, suppressing fire and managing for uniform domestic cattle grazing has been to the detriment of many wildlife species, especially grassland birds as a result of the homogenization of vegetation structure and composition⁹.

PBG uses prescribed burning a patch within a fenced pasture and free access grazing allows bison and cattle to choose burned or unburned areas and moves fire and grazing disturbances around the landscape causing vegetation patterns to shift through space and time^{8,14}. Depending upon the elapsed time-since-fire of a given patch, the probability of grazing or burning of that patch varies due to the resulting vegetation structure^{11,15,16}. Moreover, PBG is different from conventional approaches to grazing management because it attempts to integrate conservation of biodiversity with livestock production which is becoming increasingly important in many rangeland landscapes. PBG also differs because it does not require additional fencing to manipulate grazer movements which is common amongst other grazing management practices.

The impetus for PBG research was out of concern for natural resource conservation, with a major focus on wildlife population ecology, native plant conservation and soil processes, etc. More recently, implications for sustaining beef cattle production have been emerging in the literature. Our objectives for this review were to

examine the literature for effects of PBG in two areas: (1) livestock production and inputs, and (2) maintaining or improving ecosystem function and biodiversity as necessary for supporting sustainable livestock production. We place this information within the context of North America relative to the promotion of PBG as an alternative for livestock management. We also identify gaps in the knowledge base and recommend areas for additional study.

Materials and Methods

For this review, we defined PBG as the free interaction of native wildlife and domestic livestock with burned and unburned areas on the landscape through space and time. By assessing studies that considered native wildlife, we identified potential benefits to livestock. We defined PBG as the applied managerial approach to restore the interaction of fire and grazing as an ecological process that uses pyric-herbivory or fire-driven grazing¹¹. We searched the literature using Google Scholar and Web of Science academic search platforms for the following terms: ‘patch-burn grazing’, ‘fire–grazing interaction’, ‘pyric herbivory’ and combinations of ‘fire’, ‘grazing’, ‘livestock’, ‘wildlife’, ‘plants’, ‘birds’, ‘composition’ and ‘structure’. In a few instances, we included information from the non-technical literature including extension bulletins, theses/dissertations and agency reports. We recognized that such information has not always been vetted by the peer-review process, but since these reports may be the only results available regionally, they provided a perspective of where research has occurred and provided ideas for additional empirical inquiry (Fig. 2).

To understand the effects of PBG across environmental gradients, we then examined the identified literature for consistent livestock-production response variables suitable for meta-analyses and the calculation of effect sizes. Seven studies had consistent measures of cattle weight gains (i.e., calf weaning weight or stocker cattle gains) and three studies had consistent comparative measures of forage quality in burned and unburned areas. Only two studies had consistent resource selection functions for fire and were not suitable for meta-analytic statistics. We calculated the effect size of calf weight gains, yearling weight gains compared with not burning, yearling weight gains compared with burning, and forage quality data using an estimate of the standard mean difference for a measure of effect size using Hedges’ *d*. Hedges’ *d* is more suitable for unequal sampling variances in the experimental and control groups than Cohen’s *d* and accounts for small sample sizes with a correction term^{17–19}. Effect sizes were compared by assessing the variance in the effect and relative magnitude.

We begin by reviewing the effects of PBG on livestock production and inputs and then move to maintaining ecosystem function and biodiversity. We present the results of

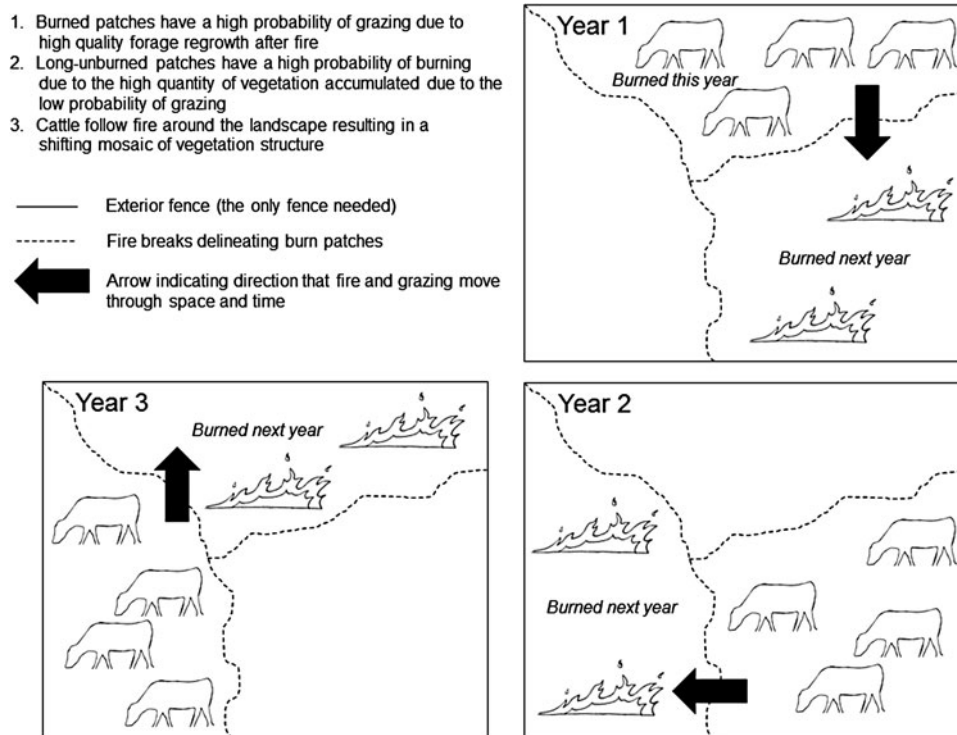


Figure 1. Functional diagram of pyric-herbivory using a 3 yr fire return interval (it does not have to be 3 yr and would likely to be variable depending on site productivity and vegetation). Note the movement of fire and grazing through space and time as cattle follow fire to the most recently burned patches.

effect size calculations and then summarize constraints and limitations to PBG knowledge currently.

Livestock Production and Inputs

Feed costs

The largest input cost of cattle production is supplemental feed, which exceeds more than half of the direct cost in cow-calf operations but is less in stocker operations²⁰. Feeding strategies attempt to overcome seasonal periods of inadequate forage quality such as the winter in perennial C4 grasslands of North America and/or periods of inadequate forage quantity such as periods of drought²¹. Reports on the value of patchy fires for cattle production in native fire-dependent ecosystems date back to the 1960s. A study in native longleaf pine—bluestem rangeland reported that patchy fires every 3 yr increased forage palatability, nutritive value, herbaceous plant dominance and cow and calf weight gains²². Patchy fires also increased cattle gains, crude protein content of forage plants and utilization of wiregrasses (*Aristida* spp. and *Sporobolus* spp.)²³. In coastal prairies, patchy fires in gulf cordgrass (*Spartina spartinae* (Trin.) Merr. ex Hitchc.) increased dietary crude protein content and *in vitro* organic matter digestibility sustaining or increasing steer gains²⁴. Forage quality of burned patches in tallgrass prairie exceeded unburned patches by a factor of four with 18 and 4% crude protein, respectively²⁵. Accordingly,

cattle use of the recently burned patch is greatly disproportionate to the area of the patch. For example, 75% of grazing time has been in the most recently burned patch in tallgrass prairie¹⁵.

PBG optimizes forage quantity in patches that have not been burned for an extended period of time – and subsequently have not been grazed – and have accumulated forage that could be considered as stockpiled forage or standing hay²⁶. Late winter fires in shortgrass steppe did not affect herbaceous plant production but did increase *in vitro* dry matter digestibility of the dominant C4 grass [*Bouteloua gracilis* (Willd. ex Kunth) Lag. ex Griffiths], providing a neutral effect on forage quantity and a short-term enhancement of forage quality²⁷. In addition, some studies have reported an increase in productivity for both C3 and C4 perennial grasses, specifically [*Paspalum smithii* (Rydb.) A. Love and *Schizachrium scoparium* (Michx.) Nash]²⁹. As a result, PBG can be strategically used to optimize forage quality and quantity, potentially mediating feed costs by providing both high quality forage (low quantity) and high quantity forage (low quality).

In the southern Great Plains of the USA, both cow-calf and stocker cattle enterprises reported that PBG did not decrease production and at times maximized production over multiple years, compared with the regionally common grazing practices which did not include fire or burned entire pastures every few years^{30,31}. In the northern Great Plains of the USA, PBG has maintained or

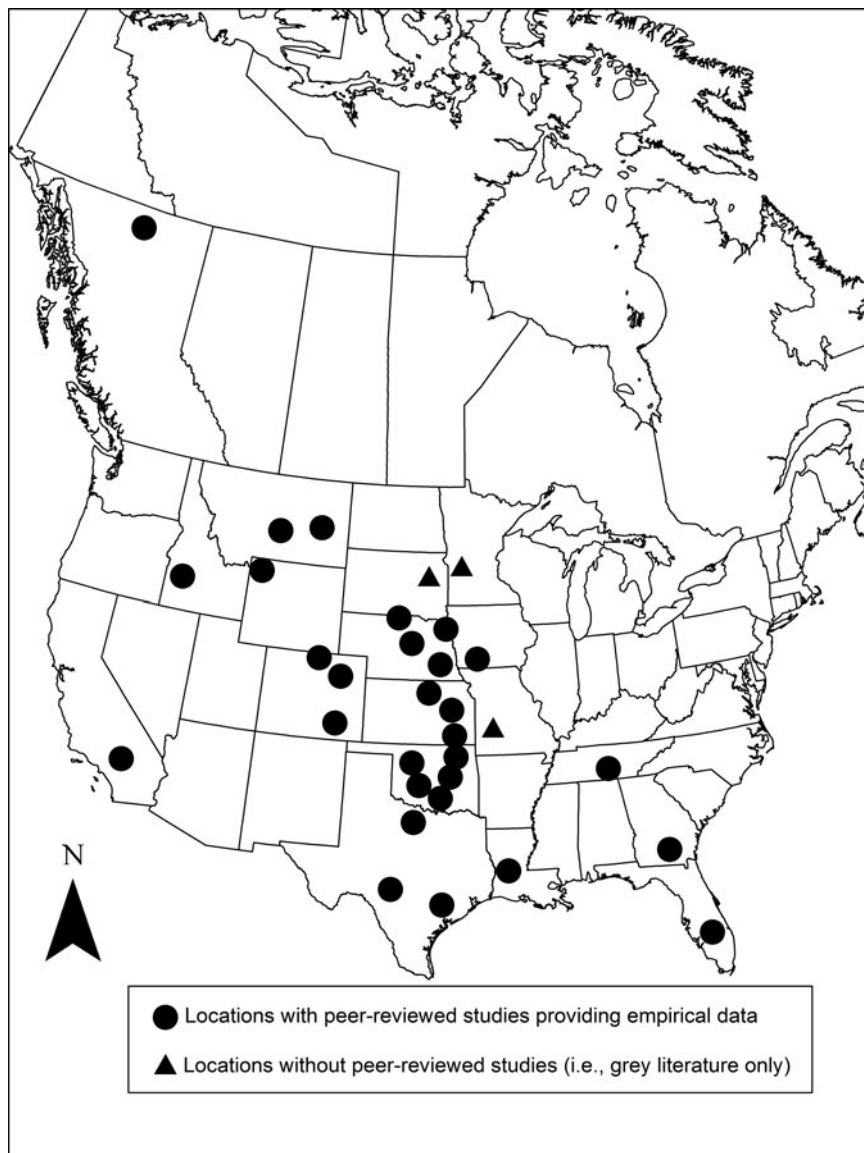


Figure 2. Study locations assessing the interaction of fire and grazing in North America.

increased weaning weights of calves and maintained body condition of mature cows^{32,33}. Managers have also reported delaying winter supplemental feeding due to the extension of higher forage quality in the fall in pastures managed with PBG³⁴. Lastly, a recent multi-year study compared PBG with the practice of annually burning the entire pasture and seasonally grazing stocker cattle, a common practice in the Flint Hills region of the USA. This study reported that PBG had nearly similar animal gain during dry years providing a risk management strategy against drought³⁵.

Parasites and disease

Parasites constitute another major source of potential economic loss and input costs for cattle enterprises. A 4 yr study comparing PBG with traditional management

significantly reduced ticks (*Amblyomma americanum* L.) on both cows and calves, regardless if the control pastures were completely burned or not burned at all³⁶. PBG also reduced horn flies (*Haematobia irritans* L.) 41% compared with no burning, reducing fly levels below the economic threshold for insecticidal treatments³⁷. Reductions in face flies (*Musca autumnalis* DeGeer) have also been reported and reductions of these flies is in part due to combustion of fecal resources, but PBG reductions of flies can be limited during drought^{38,39}. Although no studies have reported the effects of fire on cattle gastrointestinal parasites, Stone's sheep (*Ovis dalli stonei* Nelson) with access to burned areas had ~10% lungworm (*Protostrongylus* spp.) infection as those grazing unburned areas only⁴⁰.

Both horn flies and ticks serve as vectors for many diseases resulting in additional input costs for medicine and

veterinary services. Conventional insecticidal management is expensive and variable in efficacy because of rapidly developing genetic resistance^{41–44}. Ticks serve as vectors for bacterial, viral and protozoal disease agents that can also lead to paralysis, toxicosis, irritation and allergy⁴². Horn flies have been implicated in the transmission of bovine leukosis virus, helminths of the skin and more⁴⁵. Animal health related costs account for 7–13% of operating costs²⁰, so reducing ecto-parasite pressure with PBG, or any effective cultural approach for that matter, will lead to a reduction in animal health costs by reducing exposure to diseases^{36,37}. Furthermore, the lack of fire leads to the encroachment of *Juniperus virginiana* (L.) that is positively correlated with *Culex tarsalis* (Coquillett), a mosquito vectoring West Nile virus, a threat to animals and humans⁴⁶.

Physical dermatitis

Many rangeland plants have physical defense mechanisms that deter grazing including thorns and pointed leaves that can injure upon contact⁴⁷. A well-known example is prickly pear cactus (*Opuntia* spp.) which has spines that reduce forage consumption, and cause physical damage to the mouth and upper GI tract of sheep, goats and cattle^{48,49}. Fire can offer a practical and economical strategy to remove the spines and reduce contact dermatitis⁴⁹. PBG in semiarid grassland also attracted pronghorn antelope (*Antilocapra americana* Ord) at densities 7–26 times greater in spring and winter burned patches than unburned patches, resulting in a 5x increase of bitten or uprooted cactus cladodes in burn patches and a 54–71% reduction of cactus during the first year of burning⁵⁰. This reduction in cactus density attributed to the interaction of fire and pronghorn grazing was maintained for at least 6 yr after burning.

Resource selection

PBG allows grazing animals to make resource selection decisions without forcing that is applied with other grazing management strategies such as cross-fencing^{15,25}. However, cross-fencing can be useful to divide large pastures into multiple paddocks to assist in locating and managing cattle. Patchy fires increase forage utilization in burn patches compared with unburned areas but still allows animals to select locations without restricting animals to a fenced paddock¹⁴. Herding could affect resource selection decisions in a low-stress scenario but human resources are increasingly difficult to find and herding would only achieve one of the many other benefits that PBG realizes.

Diet diversity, inter-animal competition and reproduction

Dietary diversity has positive associative effects for herbivores but constructing species mixtures that complement

one another in nutrient content and secondary compounds is not well understood⁵¹. PBG allows cattle to respond to burned patches and shifts the grazing decision from the plant scale to the patch scale, so consumption of a greater variety of plants is expected. This change in dietary selection is demonstrated by studies reporting PBG causing cattle to graze plant species that they typically avoid without fire^{23,52–54}. At The Nature Conservancy's Tallgrass Prairie Preserve in Osage County, Oklahoma, USA, fire and grazing have been recoupled at the landscape scale. This has allowed fire and bison grazing to freely interact, resulting in high bison reproductive rates without nutritional supplementation¹¹. The direct benefits to animal welfare may be the least understood benefit of PBG.

Herbicides for invasive weed management

Another threat to sustainable livestock enterprises is the encroachment and dominance of unpalatable exotic or native herbaceous plants that are often combatted with herbicides. In the southern Great Plains and Midwestern USA, an exotic legume, sericea lespedeza or Chinese bushclover [*Lespedeza cuneata* (Dum.Cours.) G.Don], is a threat and management challenge to cattle producers. As *L. cuneata* invades it creates monocultures that displace native grasses, alters structure and composition of plant communities and decreases overall grazable forage⁵⁵. Ranchers and conservation organizations have reported allocating a substantial portion of their operating budget spraying for *L. cuneata*, often with only marginal success^{56,57}.

A primary mechanism facilitating *L. cuneata* invasion and dominance over native plant communities is the high tannin levels that deter grazing⁵⁸. The application of PBG overcomes the tannin grazing deterrent and increases herbivory. This slows the rate of invasion—three times slower than in traditionally managed pastures⁵⁴. Functionally, PBG results in focal grazing that begins at an early plant growth stage after fire and grazing continues to perpetuate an earlier phenological stage.

Restoring the fire disturbance alone can be applied to manage other problematic weeds on North America rangelands. For example, fire reduced broom snakeweed [*Gutierrezia sarothrae* (Pursh) Britton & Rusby], prickly pear cactus (*Opuntia polyacantha* Haw.), and purple threeawn (*Aristida purpurea* Nutt.)^{59–61}. Fire may also restore the C4 grass component in areas dominated by C3 annual grasses⁶². Fire has also been effectively restored in areas that are invaded by naturalized C3 grasses but additional information is currently lacking on if or how fire may reduce exotic C3 grasses³⁹. Ultimately, PBG and the restoration of regular fire has the potential to slow exotic or invasive plant encroachment and dominance, reduce herbicide application costs and minimize losses to the grazable forage base.

Mechanical tree and brush control management

Woody plant encroachment is another threat to livestock production. Species such as eastern redcedar (*J. virginiana*) convert open grassland to closed woodland in as little as 40 yr⁶³. Historically, fires relegated these non-sprouting and fire sensitive trees to shallow soils and topography where fire was unlikely to spread. The low growing canopy of *J. virginiana* reduces herbaceous plant production and grazing capacity^{64,65}. Other *Juniperus* species, such as *Juniperus ashei* (J. Buchholz) and *Juniperus pinchotii* (Sudw.) are similarly problematic in other regions of the USA⁶⁶.

Ranchers have applied a variety of costly and temporary mechanical brush control practices (mowing, hand cutting, bulldozing, roller chopping) but fire may be the most economical and effective for non-resprouting and resprouting woody plants^{67,68}. PBG offers a practical framework for applying regular fire to reduce the need for costly mechanical brush control costs. PBG has the potential to be more effective at reducing woody plant encroachment than complete burning pastures. The burned areas draw grazing animals from unburned areas which then can accumulate adequate fuel for the next successful fire and creates fire breaks by focusing grazing and removing fine fuels in other areas⁶⁹. This pattern of fuel accumulation driven by fire-grazing patterns enhances the potential success of prescribed fires for brush control because continuous grazing and burning pastures completely may not support the frequency of burning needed^{34,67,70}.

Encroachment by resprouting shrubs is also a concern in fire-dependent ecosystems. These shrub species are able to resprout basally and/or epicormically, and are not killed by fire. Fire, however, can alter the structure of these shrubs benefitting the herbaceous plant community important for cattle grazing⁷¹. PBG with summer fires reduced cover of honey mesquite (*Prosopis glandulosa* Torr.) and other resprouting shrubs facilitating herbaceous plant recovery⁷². Therefore, regular fire has the potential to slow the invasion of undesirable plants that can reduce forage available for cattle and offset the need for expensive and temporary mechanical brush control costs.

Nitrogen (N) availability

Net primary productivity of most terrestrial ecosystems is N limited and this leads to additional input costs for livestock production⁷³. Functionally, N is critical for plant growth and microbial breakdown of cellulosic material in the rumen of cattle⁷⁴. In tallgrass prairie, PBG enhances N availability by interactively cycling nutrients rapidly with fire followed by focal grazing⁷⁵. The authors explicitly stated this interaction between fire and grazing and the resulting increase in plant available

N may offer a strategic management approach for sustaining livestock production; likely because N content is used to calculate crude protein, the primary measure of feed quality. Furthermore, the disturbance of fire in tallgrass prairie removes litter, increasing productivity, nutrient cycling and plant available N^{76,77}. In shortgrass steppe, PBG with March burns created a pulse of N with enhanced soil N availability in June and July²⁷. Considering the different inputs managers use in the attempt to distribute/increase N across the landscape (supplemental feed high in N content, fertilizer, establish exotic legumes, etc.), the accelerated nutrient cycling associated with PBG could offset these inputs. However, the short-term N pulse post-fire needs to be understood in context with potential net volatile loss of N and subsequent N:carbon dynamics relative to ecosystem stability.

Grazing distribution

Grazing distribution continues to be a major challenge for livestock production in North America⁷⁸. Managers have used a variety of inputs to manipulate grazing distribution across the landscape, including cross-fencing, mobile feeders, low moisture blocks, herding, water and more⁷⁹. Cross-fencing, in particular, is expensive for the initial construction and the required maintenance. A 2011 study estimated the cost of construction to exceed US \$5000 per kilometer, with 8% of the initial cost needed annually for maintenance⁸⁰. PBG distributes grazing by manipulating forage quality with fire as opposed to cross-fencing, developing water, moving feeds, etc. The attraction to the recently burned areas tends to override topography, distance to water or shade even in semi-arid areas and result in cattle spending a majority of time grazing in recently burned patches^{15,25,81,82}. A 3 yr study reported economic returns from PBG on tallgrass prairie could exceed those of management intensive grazing on endophyte infected tall fescue [*Schedonorus arundinaceus* (Schreb.) Dumort., nom. cons.] pasture due to almost ten times greater input cost primarily from fencing and water development^{30,83}.

Ecological costs and risks associated with cross-fencing rangeland can affect woody plant encroachment and wildlife movements. Cross-fencing increases perches for birds and serves as a recruitment pathway for bird-dispersed seeds of woody plants especially *J. virginiana*, a major threat to North American grasslands^{84,85}. Cross-fences increase collisions of Lesser Prairie-Chicken (*Tympanuchus pallidicinctus* Ridgway) and Greater Sage-Grouse (*Centrocercus urophasianus* Bonaparte)⁸⁶ and inhibit migrating ungulates such as pronghorn antelope, which typically go through fences as opposed to jumping them⁸⁷. The use of spatially and temporally discrete fires could serve as an ecological proxy for cross-fencing while reducing overhead, financial risks and ecological risks.

Ecosystem Function and Biodiversity

Plant composition and structure

The most common approaches to cattle production either completely exclude fire or burn everything, with the former being most predominant in North America with a few exceptions such as the Flint Hills in Kansas and Oklahoma USA. These two common approaches may only benefit certain segments of the plant community. For example, the interaction of fire and grazing in PBG stimulates below- and above-ground biomass of one of the most common perennial C4 grasses in mixed and tallgrass prairies, little bluestem [*Schizachyrium scoparium* (Michx.) Nash]²⁹. The interaction of fire and grazing can also improve plant root tissue quality and initiate faster cycling of N⁸⁸. In tallgrass prairie, the interactive disturbance of fire and grazing increases plant diversity due to the release of forbs that are often inhibited by the structurally dominating tallgrass species^{14,52,53}. Thus, PBG integrates fire and grazing disturbances that optimize native grasses that are critical for ruminant livestock and can increase floristic diversity of fire-dependent ecosystems.

The primary intent of PBG has been to restore patterns of landscape heterogeneity because heterogeneity is the root of biological diversity at all levels of ecological organization and scales^{11,89}. Many studies have reported that PBG increased heterogeneity of vegetation visual obstruction, or contrast between patches, at the patch scale as opposed to methods that promote homogeneity through annual burning and grazing or not burning at all^{15,90–92}. In ecosystems with a dominant shrub component, such as sand sagebrush (*Artemisia filifolia* Torr.), PBG restored heterogeneous vegetation patterns and maintained herbaceous plant dominance and plant succession⁹⁰.

However, not all studies have resulted in the desired level of heterogeneity⁹³. Constraints to heterogeneity management include overgrazing prior to attempting to burn, exotic species and stocking rate³⁹. These constraints modify the fuel bed and limit fire spread. Furthermore, the interactive effects of fire and grazing on structural heterogeneity are scale dependent and in some areas may also be constrained by topography^{94,95}. In desert grasslands, the interaction of fire and grazing can lead to a decrease in perennial grass cover but an increase in species diversity; tradeoffs that warrant further examination⁹⁶. The lack of structural heterogeneity in highly disturbed plant communities, potentially negative effects on the plant community in arid environments and variability in plant–herbivore interactions across a gradient of precipitation and evolutionary histories continues to be a gap in the literature⁹⁷.

Soil and water resources

The shifting mosaic of vegetation patterns and attraction of animals to recently burned areas overrides other

resource selection criteria for cattle and has been hypothesized to potentially reduce animal preference for riparian areas. A study in semi-arid rangeland reported PBG led cattle to select riparian areas five times less than cattle in traditionally managed pastures, effectively reducing the impact of disturbance due to grazing⁹⁸. Given the preference of cattle for both shade and water, along with predictions for a warming climate, PBG can strategically mitigate the risk to riparian areas being overutilized and degraded⁹⁹. PBG also creates a shifting pattern of vegetation structures that varies through space and time and reduces or eliminates ‘sacrifice’ areas where animals congregate resulting in degradation⁷².

A study on PBG in coarse textured sandy soils found an increased rate of erosion on burned patches although no drifting or blowouts were observed¹⁰⁰. In the same study, when spring weather promoted early plant growth, erosion was similar between burned and unburned patches¹⁰⁰. This study also found soil water content and plant productivity were unaffected by PBG but soils in burned patches were 1–3°C warmer than unburned plots. A study on silty clay loam soils also resulted in warmer soil surface, more bare ground, less litter, greater runoff depth and greater sediment loss in recently burned patches but no difference in soil compaction, soil C, or total N¹⁰¹.

Invertebrates

The subsequent effects of the interaction of fire and grazing span many trophic levels of wildlife, including invertebrates. A mesic prairie study reported 50% greater total invertebrate biomass and greater abundance of multiple invertebrate orders in the patch that was burned and focally grazed the previous year compared with traditionally managed pastures¹⁰². A similar study in semi-arid sagebrush communities reported that Araneae needs unburned areas, Hemiptera needs burned areas, and Orthoptera equally use areas that are both burned and unburned¹⁰³.

Pollinators may also benefit from PBG as Monarch butterflies (*Danaus plexippus* L.) increased concurrently with increases in the host plant green antelopehorn milkweed (*Asclepias viridis* Walter) in patch-burned pastures that used summer fires¹⁰⁴. Other butterfly studies have reported variable responses to fire and grazing with different species having different sensitivities to elapsed time since fire and grazing^{105,106}. However, it is evident from these studies that butterflies are sensitive to changes in the herbaceous plant community. The risk of not burning at all is a potential shift to a woodland state and alternatively, burning entire areas can reduce larvae and potentially eliminate populations that inhabit isolated grassland fragments^{107,108}.

Many of the native nectar plants that pollinators depend on are forbs which increase with PBG^{14,52}.

Another example is the need to maintain native vegetation by using fire to combat cedar (*Juniperus* spp.) encroachment to conserve the federally endangered American Burying Beetle (*Nicrophorus americanus* Olivier) and other grassland obligate detritivores¹⁰⁹. The spatio-temporal interaction of fire and grazing has important implications for invertebrate biodiversity.

Grassland birds

Grassland birds have been declining over the last several decades and PBG restores structural and compositional heterogeneity to the benefit of grassland bird species^{9,89,110,111}. Increased landscape heterogeneity from PBG creates greater diversity and abundance of grassland obligate birds by offering a broader range of habitat structures that benefit all life phases and help moderate thermal extremes^{9,112}. Species reported to be declining across their historical range tend to occur at the extreme ends of the spectrum of vegetation structure; Upland sandpipers (*Bartramia longicauda* Bechstein) prefer recently burned and heavily grazed patches while Henslow's sparrows (*Ammodramus henslowii* Audubon) require patches not recently burned or grazed⁹. Similar research reported increased bird species richness and greater abundance of Horned Larks (*Eremophila alpestris* L.) in PBG pastures in comparison with control pastures³⁰. Additionally, bird demographic studies reported increased nest survival for Dickcissels (*Spiza americana* Gmelin) and Grasshopper Sparrows (*Ammodramus savannarum* Gmelin) in PBG pastures compared with pastures with homogenous vegetative structure^{113,114}. The heterogeneity created by PBG increases diversity and stability in breeding and non-breeding grassland bird communities^{115,116}.

A long-term assessment of grassland birds over two decades suggests that fire and grazing must be variable in intensity of disturbance and restore heterogeneity if grassland birds are to be conserved¹¹⁰. In the western USA, Mountain plovers (*Charadrius montanus* Townsend) are also tightly coupled with the fire-grazing disturbance that creates low statured and bare ground habitat they require⁹⁵. Patchy fires are also required by Northern Bobwhite quail (*Colinus virginianus* L.) to provide the suite of vegetation structure needed for all life phases and PBG has been suggested as the best strategy for providing this habitat mosaic^{117,118}. Finally, a patchy application of disturbance to tallgrass prairie has consistently been recommended to prevent the continued decline of Greater Prairie-Chickens throughout the Flint Hills of Kansas and Oklahoma, USA^{119,120}. These results support the role of PBG in integrating grazing and biological conservation by restoring critical disturbance processes that shape grassland environments for birds obligated to this type of habitat¹²¹. These results indicate that PBG provides an alternative to homogeneous management on rangelands or the idea of managing

towards the middle which are common practices across most rangelands in North America^{15,78}.

Mammals

PBG creates a mosaic of patches with different amounts of vegetation biomass, forage quality and structure, whereby different patches may be used differently by different wildlife species. For example, deer mice (*Peromyscus maniculatus* Wagner) were ten times more abundant on burned patches, but hispid pocket mice (*Chaetodipus hispidus* Baird) were ten times more abundant on intermediate patches. Hispid cotton rat (*Sigmodon hispidus* Say & Ord), prairie vole (*Microtus ochrogaster* Wagner) and fulvous harvest mice (*Reithrodontomys fulvescens* J. A. Allen) all dominated patches not burned in >2 yr¹²². Patchy fires that are focally grazed also influenced black-tailed prairie dogs (*Cynomys ludovicianus* Ord) with colonies expanding two times faster into burned areas compared with unburned areas in shortgrass steppe^{123,124}.

Large mammals also require diversity in habitat. White-tailed deer (*Odocoileus virginianus* Zimmermann) grazed summer burned areas with peak use occurring within the first 2 months after fire¹²⁵. In sagebrush communities, elk (*Cervus elaphus* L.) had greater herbivory of burn patches the first 2 yr after fire¹²⁶. Similar long-term effects were reported for the winter nutritional plane of *C. canadensis* and mule deer (*Odocoileus hemionus* Rafinesque) with positive associative effects lasting up to 2 yr¹²⁷. The value of burned areas may be increasingly important for winter habitat and nutrition as elk and bison used burned patches more than expected especially during mid to late winter¹²⁸. From a conservation standpoint, the use of patchy fire has also been suggested as a habitat restoration tool for bighorn sheep (*Ovis canadensis* Shaw)^{129,130}. Stone's sheep (*O. dalli stonei*) in sub-alpine and alpine ranges also benefit from patchy fires due to greater forage quantity on burned range that resulted in lower internal parasite loads and greater lamb crops than sheep on unburned range⁴⁰.

Data Analyses: Effect Size of Livestock Production Variables across Gradients

Calf weaning weight: PBG versus burning entire pasture every third year

Only two studies presented suitable data for meta-analyses of calf weaning weights under PBG management^{31,33}. Both studies compared PBG with burning entire pastures every third year and the studies were located in southeastern Nebraska, USA³³ and north-central Oklahoma, USA³¹. The Nebraska study had an effect size and variance of 0.51 ± 0.69 and the Oklahoma study had an effect size of -0.10 ± 0.50 .

Overall calf weaning weights under PBG did not differ from those pastures managed with fire every third year (effect size = 0.16 ± 0.60 ; Fig. 3A).

Yearling cattle weight gain: PBG versus burning entire pasture every third year or burning annually

Only two studies presented suitable data for meta-analyses of yearling cattle weight gains under PBG management compared with management from different spatio-temporal applications of fire^{31,35}. Both studies compared PBG with burning entire pastures every third year and both studies were located in north-central Oklahoma, USA^{31,35}. The effect size of -0.36 ± 0.51 and -0.25 ± 0.34 (Stillwater and Pawhuska, respectively) did not differ. Overall, yearling cattle weight gain under PBG did not differ from those pastures managed with fire every third year (effect size = -0.29 ± 0.42 ; Fig. 3B).

Yearling cattle weight gain: PBG versus not burning at all

Three studies presented suitable data for meta-analyses of yearling cattle weight gains under PBG management compared with not burning at all^{24,31,131}. Studies were located across an annual precipitation gradient ranging from 339 mm in the shortgrass steppe near Nunn, Colorado, USA¹³¹, 725 mm in the mixed grass prairie near Bessie, Oklahoma, USA³¹, and 877 mm in the coastal prairie near Sinton, Texas, USA²⁴. The effect size was 0.36 ± 0.51 in the shortgrass steppe location, 0.81 ± 0.20 in the mixed grass prairie location and 1.49 ± 0.86 in the coastal prairie location, with error bars that only overlapped zero in the shortgrass steppe location. Overall, yearling cattle weight gain with PBG was greater than when not burned (effect size = 0.80 ± 0.52 ; Fig. 3C). Furthermore, this limited data set suggests the possibility for increasing yearling cattle weight gains as annual precipitation increases with patchy fires.

Forage quality

Three studies presented suitable data for meta-analyses of forage quality with and without fire^{25,98,131}. Studies were located across an annual precipitation gradient ranging from 339 mm in the shortgrass steppe near Nunn, Colorado, USA¹³¹, 725 mm in the mixed grass prairie near Bessie, Oklahoma, USA⁹⁸, and 1,005 mm in the tallgrass prairie near Pawhuska, Oklahoma, USA²⁵. Crude protein was $15.5\% \pm 0.8$ in burned patches and $8.8\% \pm 0.8$ in unburned patches in the shortgrass steppe location, $15.5\% \pm 0.3$ and $7.6\% \pm 0.2$, respectively, in the mixed grass prairie location, and $16.9\% \pm 0.5$ and $4.1\% \pm 0.1$, respectively, in the tallgrass prairie location (Fig. 4A). The effect size was 3.8 ± 1.4 in the shortgrass steppe location, 5.5 ± 1.4 in the mixed grass prairie location and 12.5 ± 6.9

in the tallgrass prairie location, with error bars that never overlapped zero (Fig. 4B).

Overall forage quality with fire was greater than forage quality without fire (effect size = 5.4 ± 3.2 ; error bars did not overlap zero; Fig. 4B). Furthermore, this limited data set suggests the strength of attraction to burned areas increases with precipitation. The up-side potential of forage quality post-fire only had a range of 1.4% but the down-side consequence of forage quality without fire had a range of 4.7%, indicating that as you move across the precipitation gradient the attraction to burned areas may be greater in higher precipitation zones due to a greater feedback, and potential negative consequence, driven by low-forage quality in unburned areas. This corresponds to 75% of grazing time spent in burned patches in tallgrass prairie¹⁵ compared with only 31% in mixed grass prairie¹³¹.

Discussion and Conclusions

Discussion

This review has examined PBG as a livestock management alternative for fire-prone ecosystems in North America, a unique approach to grazing management on this continent with potential international application^{2,4-7,133-135}. This review has been restricted to a single continent, and the majority of recent research has come from the Great Plains of North America. However, empirical studies span temperature and precipitation gradients (Fig. 2). The literature supports PBG as an alternative management strategy to sustain production by sustaining or optimizing cattle gains, optimizing forage quality and quantity, mitigating the negative effects of drought, reducing parasite pressure and insecticide treatments, reducing chemical and mechanical weed and brush control inputs, reducing N additions and offering an alternative to expensive cross-fencing and water development to overcome grazing distribution constraints (Table 1). Globally, low-input pasture based livestock production systems are essential for meeting societal demands for goods and services but additional strategies that potentially mitigate climate and market fluctuations will enhance sustainability^{136,137}. Because input costs and drought threaten the sustainability of livestock production, realizing the potential benefits of PBG to offset these threats is a potential sustainability strategy for fire-prone ecosystems of North America¹³⁸.

A critical benefit of the PBG process driven approach is the ability to integrate livestock production and natural resource conservation in multifunctional working landscapes by restoring critical ecological functions and maintaining perennial herbaceous vegetation; features that should be considered part of a renewable agriculture and food system¹³⁹⁻¹⁴¹ (Table 2). Land managers should not have to choose one over the other but rather should

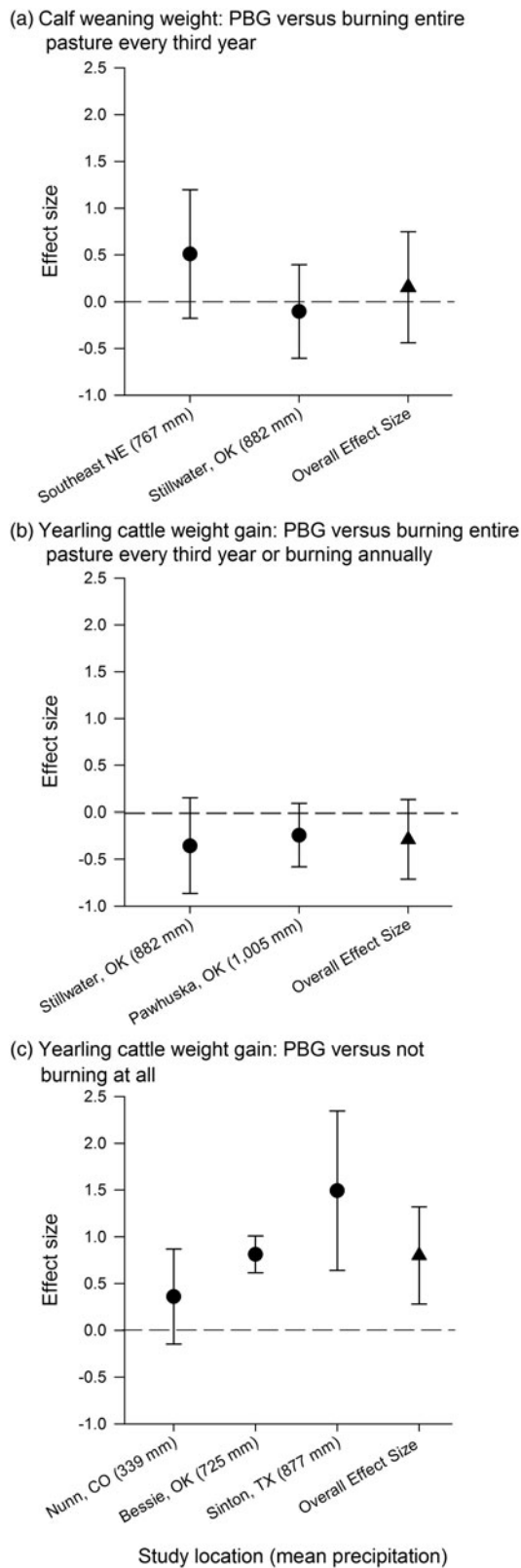


Figure 3. Patch-burn grazing (PBG) effect sizes along precipitation gradients for (A) calf weaning weights versus burning entire pastures every third year, (B) yearling cattle weight gains versus burning entire pastures every third year or burning annually, and (C) yearling cattle weight gains versus not burning at all.

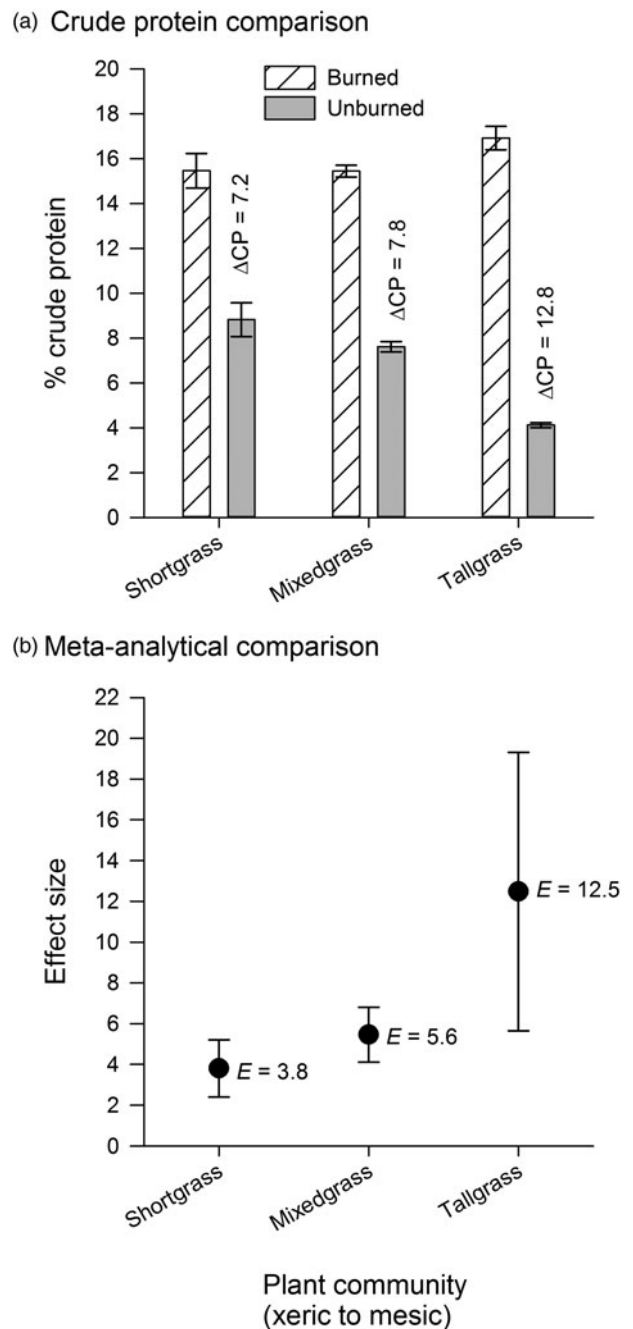


Figure 4. Crude protein comparison along the precipitation gradient for burned and unburned sites. (A) Presents the mean and (B) presents the effect size.

be able to integrate the two in a complementary approach. Many of the native insects, birds and mammals are dependent on fire-grazing processes to increase suitability for breeding habitats, thermal regulation and foraging opportunities^{106,113,122}. The resulting patterns and vegetation succession optimize the variable habitat and foraging needs of a wide spectrum of species⁸⁹. Furthermore, native plant species and communities are maintained and woody plant encroachment is minimized. Native wildlife species that are of concern can be

Table 1. Summary of potential benefits of patch-burn grazing (PBG) for animal production.

Benefit	Effect (US state location)	Source
Direct production benefits		
Feed costs	Optimized forage quality and quantity (OK) Maintained quantity and enhanced digestibility (CO) Improved forage quality and creates grass bank (TN) Increased productivity of <i>Pascopyrum smithii</i> (MT) Increased productivity of <i>Schizachyrium scoparium</i> (OK) Delaying winter feeding (OK)	Allred <i>et al.</i> ²⁵ Augustine <i>et al.</i> ²⁷ McGranahan <i>et al.</i> ⁹² Vermeire <i>et al.</i> ²⁸ Limb <i>et al.</i> ²⁹ Weir <i>et al.</i> ³⁴
Cow-calf production	Sustained cow body condition and calf gains (OK) Sustained cow body condition and calf gains (NE) Increased body condition and calf gains (MN, ND) Increased cow and calf weight gains (LA) Increased cow and calf weight gains (GA)	Limb <i>et al.</i> ³¹ Winter <i>et al.</i> ³³ Baumann ³² Duvall & Whitaker ²² Hilmon & Hughes ²³
Stocker cattle production	Sustained gains of stocker cattle (OK)	Fuhlendorf & Engle ¹⁵
	Sustained or increased gains of stocker cattle (OK) Sustained or increased gains of stocker cattle (CO) Sustained or increased gains of stocker cattle (MO) Sustained or increased gains of stocker cattle (OK) Sustained or increased gains of stocker cattle (TX)	Limb <i>et al.</i> ³¹ Augustine & Derner ¹³¹ Jamison & Underwood ³⁰ Allred <i>et al.</i> ³⁵ Angell <i>et al.</i> ²⁴
Optimize reproduction	High bison reproductive rates without supplement (OK)	Fuhlendorf <i>et al.</i> ¹¹
Drought losses	Stabilized gains versus burn everything (OK)	Allred <i>et al.</i> ³⁵
Parasites	41% horn fly reduction versus no fire (OK, IA) Reduced horn fly and face fly but drought can inhibit (OK, IA) 57% tick reduction versus burn all or no fire (OK) 4–10x lower GI parasites in wild sheep (BC Canada)	Scasta <i>et al.</i> ³⁷ Scasta <i>et al.</i> ³⁸ Polito <i>et al.</i> ³⁶ Seip and Bunnell ⁴⁰
Disease exposure	Reduces exposure to disease vectoring insects	(see parasites above)
Thermal regulation	Optimizes options for thermal regulation (OK)	Allred <i>et al.</i> ⁹⁹
Physical dermatitis	Fire can reduce plant structures damaging mouth (TX)	McMillan <i>et al.</i> ⁴⁹ ; Migaki <i>et al.</i> ⁴⁸
	Pronghorn reduced cactus density in patch burns (CO) Altered plant selection in burned patch (OK) Increased utilization of broadleaf plants (NE) Increased utilization of <i>Sericea lespedeza</i> (OK)	Augustine & Derner ⁵⁰ Coppedge <i>et al.</i> ⁵² Helzer & Steuter ⁵³ Cummings <i>et al.</i> ⁵⁴
Diet diversity		
Indirect production benefits		
Herbicide inputs	3x slower invasion <i>Sericea lespedeza</i> versus no fire (OK) Reduced broom snakeweed and prickly pear (CO) Reduced broom snakeweed (NM) Reduced purple threeawn (MT) Reduced cactus density in patch burns Fire in mesic grassland for resprouting shrub control (KS) Fire in semi-arid range for resprouting shrub control (TX)	Cummings <i>et al.</i> ⁵⁴ Augustine & Michunas ⁶⁰ McDaniel <i>et al.</i> ⁵⁹ Strong <i>et al.</i> ⁶¹ Augustine & Derner ⁵⁰ Heisler <i>et al.</i> ⁷¹ Teague <i>et al.</i> ⁷²
Mechanical brush control	Mechanical 2–5x more expensive for <i>Juniperus</i> (OK) Fire in mesic sagebrush steppe for <i>Juniperus</i> control (ID) Fire in tallgrass prairie for <i>Juniperus</i> control (OK) Fire in mixedgrass prairie for <i>Juniperus</i> control (TX)	Bidwell <i>et al.</i> ⁶⁸ Clark <i>et al.</i> ⁸¹ Limb <i>et al.</i> ⁶⁵ Ansley <i>et al.</i> ⁶⁶
Nitrogen (N)	Enhances N availability in burn patch (OK) Enhances N availability in in semi-arid ecosystems (CO)	Anderson <i>et al.</i> ⁷⁵ Augustine <i>et al.</i> ²⁷
Grazing distribution	Cattle strongly attracted to burned areas; effective and inexpensive grazing distribution tool (OK) Cattle grazed 75% of time in burned area (OK) Overcomes distribution constraints w/o fence (MO) Improved selection of previously unused areas (ID) Selected burned areas with precipitation pulses (CO)	Vermeire <i>et al.</i> ¹⁴ Fuhlendorf & Engle ¹⁵ Davit & Alleger ⁸³ Clark <i>et al.</i> ⁸¹ Augustine & Derner ¹³¹

Table 2. Summary of potential ecological benefits associated with the interaction of fire and grazing applied as patch-burn grazing (PBG).

Ecological benefit	Sources
Plant composition and structure	
Regulate woody plant encroachment/dominance	Bidwell et al. ⁶⁸ ; Kerby et al. ⁶⁹ ; Teague et al. ⁷² ; Winter et al. ⁹⁰ ; Weir et al. ³⁴
Increase plant diversity	Vermeire et al. ¹⁴ ; Coppedge et al. ⁵²
Optimize vegetation heterogeneity	Fuhlendorf and Engle ¹⁵ ; Winter et al. ⁹⁰ ; Leis et al. ⁹¹ ; McGranahan et al. ²⁶
Stimulates above/below ground biomass C4 grasses	Limb et al. ²⁹
Increased plant root tissue quality and nutrient cycling	Johnson & Matchett ⁸⁸
Removes detritus increasing productivity	Knapp & Seastedt ⁷⁶ ; Anderson et al. ⁷⁵
Soil water resources	
Reduces animal preference/use of riparian areas	Allred et al. ⁹⁹
Reduces degradation of sacrifice areas	Teague et al. ⁷²
Wildfire risk	
Minimize the spread of wildfires and potentially increase ability to absorb wildfires to protect fire-sensitive areas from catastrophic fires	Kerby et al. ⁶⁹
Invertebrates	
Increase invertebrate diversity and abundance	Engle et al. ¹⁰² ; Doxon et al. ¹⁰³
Benefit pollinators	Vogel et al. ¹⁰⁸ ; Baum and Sharber ¹⁰⁴
Benefit detritivores	Walker & Hoback ¹⁰⁹
Mosquitoes vectoring West Nile virus (a threat to some species) prefer woody invaded areas	O'Brien & Reiskind ¹⁴⁶
Grassland birds	
Increased diversity and stability	Fuhlendorf et al. 2006 ⁹ ; Powell ¹¹⁰ ; Jamison & Underwood ³⁰ ; Coppedge et al. ¹¹² ; Augustine & Derner ⁹⁵
Increased survival	Churchwell et al. ¹¹³ ; Hovick et al. ¹¹⁵ ; Hovick et al. ¹¹⁶ ; Hovick et al. ¹¹⁶ ; McNew et al. ¹²⁰
Increased nesting cover	
Small mammals	
Optimize habitat benefiting composition	Fuhlendorf et al. ¹²²
Prairie dogs expand more rapidly in burned areas	Augustine et al. ¹²³ ; Breland ¹²⁴
Large mammals	
White-tailed deer increased use 2 months after fire	Meek et al. ¹²⁵
White-tailed and mule deer winter nutrition improved	Hobbs & Spowart ¹²⁷
Pronghorn density higher in burned patches	Augustine & Derner ¹³¹
Elk increased use for up to 2 yr after fire	Pearson et al. ¹²⁸ ; Dyke and Darragh ¹²⁷
May be used to restore bighorn sheep habitat	Bleich et al. ¹²⁹ ; Holl et al. ¹³⁰
Increased lamb crop of Stone's sheep	Seip & Bunnell ⁴⁰

managed in concert with cattle production and potentially enhanced¹⁴².

Potential negative impacts and limitations

The application of PBG is not without limitations or knowledge gaps. Some ecosystems may be so constrained by moisture that fire did not occur very often and large ungulate grazing was not a prevalent disturbance; for example, the four major North American deserts, and some ecosystems, such as the Palouse prairie⁷⁸ (Fig. 2). Some ecosystems have fire sensitive species that are critical to conservation, such as big sagebrush communities (*Artemisia tridentata* Nutt.) and patchy fires may need to be reconsidered and modified in terms of spatial scale, temporal scale and seasonality¹³⁶. However,

sagebrush communities are threatened not only by the encroachment of woody shrubs such as *Juniperus* spp. and *Pinus* spp., but are also threatened by wildfire that could cause rapid and expansive mortality due to the intense fire behavior and extent¹³⁷. Thus, additional research is needed on how patchy fires can be applied in a sustainable manner to optimize non-sprouting *Artemisia* spp. and minimize woody plant encroachment and wildfire threats. Regarding semi-arid rangelands, a recent PBG study using a 4 yr fire return interval reported grazing preference for the most recently burn patch, however, the preference was lower than the reports from mesic ecosystems¹³¹. Another study in sagebrush steppe showed that patchy fires altered cattle resource selection and overcame limitations of slope, sagebrush dominance and distance to upland water and lasted for 2–3 yr⁸¹. Drought

also has confounded the effects of PBG on grassland birds and livestock parasites and if climate forecasts continue, additional research on the efficacy under precipitation extremes is needed^{132,143}. Therefore, we suggest additional research is needed on the controls of the strength and timing of the fire–grazing interaction in more arid ecosystems and how to apply prescribed fire to mimic historical fire return intervals across broad temperature and precipitation gradients. This type of information will broaden our understanding of how herbivores respond to fire across a precipitation gradient and assist managers in tailoring the spatial and temporal prescription for fire–grazing interactions accordingly. Further research is also needed on the additional cost-benefit analyses of diet optimization, the livestock effects from burning in different seasons of the year and potential effects on gastrointestinal parasites of livestock. Lastly, while not all studies have resulted in the desired level of structural heterogeneity, the potential constraint of grazing management decisions and the time required for fire to effectively drive grazing patterns needs additional research^{39,93}.

Conclusions

Ultimately, the data from numerous studies are the evidence that PBG can benefit cattle production, ecosystem function and rural citizens over the long-term and is a renewable livestock system in fire prone plant communities. The broad geographical range of studies in North America indicates that the attraction of herbivores to recently burned patches spans both precipitation and temperature gradients (Fig. 2). While some of the studies discussed here are not strictly PBG with livestock, they do express the need for frequent fire to control woody plants. Consequently, fire is critical and PBG is a method that re-incorporates fire while integrating grazing. PBG is a bottom-up approach to grazing management that is ecologically process-based and low-input allowing animals to behave and respond to heterogeneity. Conversely, most other grazing strategies, such as rotational grazing, are top-down approaches that impose command-and-control and are high-input^{144,145}. High-input command-and-control approaches do not always result in increases in production and often result in lower animal performance and lower long-term sustainability⁸⁵.

Perhaps the greatest importance of PBG to livestock production and fire management is using patchy fires to drive grazing and vegetation patterns to overcome the forage versus fuel paradox in a fenced off landscape. In other words, when entire pastures are burned, all forage was consumed as fuel by fire and livestock have low forage availability until adequate moisture is available³⁵. PBG overcomes that relationship by optimizing fuel accumulation in unburned patches to increase fire intensity and mortality on woody plants or serve as a forage reserve during drought. PBG integrates fire and grazing

without having to sacrifice one or the other; deferring grazing to accumulate enough fuel to burn and to woody plant invasion or grazing and not being able to burn at all.

Restoration of fire and grazing with PBG fundamentally embraces variation through space and time, a diametric opposite to the utilitarian model of uniform utilization promoted by conventional management¹⁴⁶. Embracing variation and disturbances has been suggested to increase resilience and sustainability of livestock production systems¹⁴⁷. The value of PBG for sustaining ecosystem goods and services is clearly evident in the studies evaluated in this review and is driving its application beyond the core area of the Great Plains where the majority of research has been conducted. With fragmentation of the landscape, woody plants often encroach and alter herbaceous plant communities. Restoring fire to portions of these landscapes has the potential to mitigate many of the unintended consequences of fire suppression⁸⁸.

Acknowledgements. The authors thank many researchers who have developed the foundation of solid research that this review stands on. This review and meta-analysis was supported by the Iowa Department of Natural Resources, Iowa Agricultural and Home Economics Experiment Station, Oklahoma Agricultural Experiment Station, the Iowa State Wildlife Grants program grant #-U-2-R-1 in cooperation with the U.S. Fish and Wildlife Service, Wildlife and Sport Fish Restoration Program (#-U-2-R-1), and the University of Wyoming Department of Ecosystem Science and Management within the College of Agriculture and Natural Resources.

References

- 1 Bond, W.J. and Keeley, J.E. 2005. Fire as a global 'herbivore': The ecology and evolution of flammable ecosystems. *Trends in Ecology and Evolution* 20(7):387–394.
- 2 Yadava, P.S. 1990. Savannas of North-east India. *Journal of Biogeography* 17:385–394.
- 3 Vinton, M.A., Hartnett, D.C., Finck, E.J., and Briggs, J.M. 1993. Interactive effects of fire, bison (*Bison bison*) grazing and plant community composition in tallgrass prairie. *American Midland Naturalist* 129(1):10–18.
- 4 Kramer, K., Groen, T.A., and van Wieren, S.V. 2003. The interacting effects of ungulates and fire on forest dynamics: An analysis using the model FORSPACE. *Forest Ecology and Management* 181(1):205–222.
- 5 Archibald, A.S., Bond, W.J., Stock, W.D., and Fairbanks, D. H.K. 2005. Shaping the landscape: Fire-grazer interactions in an African savanna. *Ecological Applications* 15:96–109.
- 6 Anderson, R.C. 2006. Evolution and origin of the Central Grassland of North America: Climate, fire, and mammalian grazers. *Journal of the Torrey Botanical Society* 133(4):626–647.
- 7 Murphy, B.P. and Bowman, D.M. 2007. The interdependence of fire, grass, kangaroos and Australian Aborigines: A case study from central Arnhem Land, northern Australia. *Journal of Biogeography* 34(2):237–250.

- 8 Fuhlendorf, S.D., Engle, D.M., Kerby, J., and Hamilton, R. 2009. Pyric herbivory: Rewilding landscapes through the recoupling of fire and grazing. *Conservation Biology* 23(3):588–598.
- 9 Fuhlendorf, S.D., Harrell, W.C., Engle, D.M., Hamilton, R. G., Davis, C.A., and Leslie, D.M. 2006. Should heterogeneity be the basis for conservation? Grassland bird response to fire and grazing. *Ecological Applications* 16(5):1706–1716.
- 10 Biondini, M.E., Steuter, A.A., and Hamilton, R.G. 1999. Bison use of fire-managed remnant prairies. *Journal of Range Management* 52:454–461.
- 11 Fuhlendorf, S.D. and Engle, D.M. 2001. Restoring heterogeneity on rangelands: Ecosystem management based on evolutionary grazing patterns. *Bioscience* 51(8):625–632.
- 12 Hamilton, R.G. 2007. Restoring heterogeneity on the tall-grass prairie preserve: Applying the fire-grazing interaction model. In R.E. Masters and K.E.M. Galley (eds). 23rd Tall Timbers Fire Ecology Conference: Fire in Grassland and Shrubland Ecosystems. Tall Timbers Research Station, Tallahassee, FL. p. 163–169.
- 13 Pyne, S.J. 1997. *America's Fires*. Forest History Society, Durham, NC.
- 14 Vermeire, L.T., Mitchell, R.B., Fuhlendorf, S.D., and Gillen, R.L. 2004. Patch burning effects on grazing distribution. *Rangeland Ecology and Management* 57(3):248–252.
- 15 Fuhlendorf, S.D. and Engle, D.M. 2004. Application of the fire-grazing interaction to restore a shifting mosaic on tall-grass prairie. *Journal of Applied Ecology* 41(4):604–614.
- 16 Wiens, J.A. 2000. Ecological heterogeneity: An ontogeny of concepts and approaches. In M.J. Hutchings, E.A. Juhn, and A.J.A. Stewart (eds). *The Ecological Consequences of Environmental Heterogeneity*. Blackwell Science, Hoboken, NJ. p. 452.
- 17 Gurevitch, J. and Hedges, L.V. 1993. Meta-analysis: Combining the results of independent experiments. In S. M. Scheiner and J. Gurevitch (eds). *Design and Analysis of Ecological Experiments*. Chapman & Hall, New York, USA. p. 378–398.
- 18 Osenberg, C.W., Sarnelle, O., and Cooper, S.D. 1997. Effect size in ecological experiments: The application of biological models in meta-analysis. *American Naturalist* 150: 798–812.
- 19 Rosenberg, M.S., Adams, D.C. and Gurevitch, J. 1999. *MetaWin: Statistical Software for Meta-analysis*. Version 2.0. Sinauer, Sunderland, MA, USA. p. 133.
- 20 Short, S.D. 2001. Characteristics and production costs of US cow-calf operations. United States Department of Agriculture. Economic Research Service. Statistical Bulletin Number 974-3.
- 21 Waterman, R.C., Geary, T.W., Paterson, J.A., and Lipsey, R.J. 2012. Early weaning in Northern Great Plains beef cattle production systems: I. Performance and reproductive response in range beef cows. *Livestock Science* 148(1):26–35.
- 22 Duvall, V.L. and Whitaker, L.B. 1964. Rotation burning: A forage management system for longleaf pine-bluestem ranges. *Journal of Range Management* 17:322–326.
- 23 Hilmon, J.B. and Hughes, R.H. 1965. Fire and forage in the wiregrass type. *Journal of Range Management* 18(5): 251–254.
- 24 Angell, R.F., Stuth, J.W., and Drawe, D.L. 1986. Diets and liveweight changes of cattle grazing fall burned gulf cordgrass. *Journal of Range Management* 39(3):233–236.
- 25 Allred, B.W., Fuhlendorf, S.D., Engle, D.M., and Elmore, R. D. 2011. Ungulate preference for burned patches reveals strength of fire-grazing interaction. *Ecology and Evolution* 1(2):132–144.
- 26 McGranahan, D.A., Henderson, C.B., Hill, J.S., Raicovich, G.M., Wilson, W.N., and Smith, C.K. 2014. Patch-burning improves forage quality, creates grass-bank in old-field pasture: Results of a demonstration trial. *Southeastern Naturalist* 13(2):200–207.
- 27 Augustine, D.J., Derner, J.D., and Milchunas, D.G. 2010. Prescribed fire, grazing, and herbaceous plant production in shortgrass steppe. *Rangeland Ecology and Management* 63(3):317–323.
- 28 Vermeire, L.T., Crowder, J.L., and Wester, D.B. 2014. Semiarid rangeland is resilient to summer fire and postfire grazing utilization. *Rangeland Ecology and Management* 67:52–60.
- 29 Limb, R.F., Fuhlendorf, S.D., Engle, D.M., and Kerby, J.D. 2011a. Growing-season disturbance in tallgrass prairie: Evaluating fire and grazing on *Schizachyrium scoparium*. *Rangeland Ecology and Management* 64(1):28–36.
- 30 Jamison, B. and Underwood, M. 2008. Evaluation of a grazing system for maintaining grassland integrity and improving upland bird habitat. Missouri Department of Conservation Final Project Report. Available at Web site http://www.nrcs.usda.gov/Internet/FSE_DOCUMENTS/stelprdb1044913.pdf (verified 5 December 2013).
- 31 Limb, R.F., Fuhlendorf, S.D., Engle, D.M., Weir, J.R., Elmore, R.D., and Bidwell, T.G. 2011b. Pyric-herbivory and cattle performance in grassland ecosystems. *Rangeland Ecology and Management* 64:659–663.
- 32 Baumann, P. 2013. Prescribed fire can yield positive results, but may not be for everyone. Grassroots, South Dakota Grassland Coalition. 15, 1–2. Available at Web site http://www.sdgrass.org/uploads/1/8/6/5/18654664/may_2013_news_letter.pdf (verified 8 December 2013).
- 33 Winter, S.L., Fuhlendorf, S.D., and Goes, M. 2014. Patch-burn grazing effects on cattle performance: Research conducted in a working landscape. *Rangelands* 36(3):2–7.
- 34 Weir, J.R., Fuhlendorf, S.D., Engle, D.M., Bidwell, T.G., Chad Cummings, D., Elmore, D., Limb, R.F., Allred, B.W., Scasta, J.D., and Winter, S.L. 2013. Patch Burning: Integrating Fire and Grazing to Promote Heterogeneity. E-998. Oklahoma Cooperative Extension Service, Oklahoma State University, Stillwater, OK. p. 36.
- 35 Allred, B.W., Scasta, J.D., Hovick, T.J., Fuhlendorf, S.D., and Hamilton, R.G. 2014. Spatial heterogeneity stabilizes livestock productivity in a changing climate. *Agriculture, Ecosystems and Environment* 193:37–41.
- 36 Polito, V.J., Baum, K.A., Payton, M.E., Little, S.E., Fuhlendorf, S.D., and Reichard, M.V. 2013. Tick abundance and levels of infestation on cattle in response to patch burning. *Rangeland Ecology and Management* 66(5): 545–552.
- 37 Scasta, J.D., Engle, D.M., Talley, J.L., Weir, J.R., Stansberry, J.C., Fuhlendorf, S.D., and Harr, R.N. 2012. Pyric-herbivory to manage horn flies (Diptera: Muscidae) on cattle. *Southwestern Entomologist* 37(3):325–334.

- 38 Scasta, J.D., Weir, J.R., Engle, D.M., and Carlson, J.D. 2014. Combustion of cattle fecal pats ignited by prescribed fire. *Rangeland Ecology and Management* 67(2):229–233.
- 39 Scasta, J.D. 2014. Implications of pyric-herbivory on central North American grassland ecology, management and production. Oklahoma State University Dissertation, Stillwater, OK.
- 40 Seip, D.R. and Bunnell, F.L. 1985. Nutrition of Stone's sheep on burned and unburned ranges. *Journal of Wildlife Management* 49:397–405.
- 41 Birkett, M.A., Agelopoulos, N., Jensen, K., Jespersen, J.B., Pickett, J.A., Puijs, H.J., Thomas, G., Trapman, J.J., Wadhams, L.J., and Woodcock, C.M. 2004. The role of volatile semiochemicals in mediating host location and selection by nuisance and disease-transmitting cattle flies. *Medical and Veterinary Entomology* 18(4): 313–322.
- 42 de la Fuente, J., Almazan, C., Canales, M., Perez de la Lastra, J.M., Kocan, K.M., and Willadsen, P. 2007. A ten-year review of commercial vaccine performance for control of tick infestations on cattle. *Animal Health Research Reviews* 8:23–28.
- 43 Jongejan, F. and Uilenberg, G. 2004. The global importance of ticks. *Parasitology* 129:S3–S14.
- 44 Oyarzún, M.P., Quiroz, A., and Birkett, M.A. 2008. Insecticide resistance in the horn fly: Alternative control strategies. *Medical and Veterinary Entomology* 22: 188–202.
- 45 Buxton, B.A., Hinkle, N.C., and Schultz, R.D. 1985. Role of insects in the transmission of bovine leukosis virus: Potential for transmission by stable flies, horn flies, and tabanids. *American Journal of Veterinary Research* 46(1): 123–126.
- 46 O'Brien, V.A. and Reiskind, M.H. 2013. Host-seeking mosquito distribution in habitat mosaics of southern Great Plains cross-timbers. *Journal of Medical Entomology* 50(6):1231–1239.
- 47 dos Reis, V.M. 2010. Dermatitis due to plants (phyto-dermatosis). *Anais Brasileiros de Dermatologia* 85(4): 479–489.
- 48 Migaki, G., Hinson, L.E., Imes, G.D., Jr, and Garner, F.M. 1969. Cactus spines in tongues of slaughtered cattle. *Journal of the American Veterinary Medicine Association* 155(9):1489–1492.
- 49 McMillan, Z., Scott, C.B., Taylor, C.A., Jr, and Huston, J.E. 2002. Nutritional value and intake of prickly pear by goats. *Journal of Range Management* 55(2):139–143.
- 50 Augustine, D.J. and Derner, J.D. 2015. Patch burn grazing management in semiarid grassland: Consequences for pronghorn, Plains prickly pear and wind erosion. *Rangeland Ecology and Management* 68(1):40–47.
- 51 Provenza, F.D., Villalba, J.J., Haskell, J., MacAdam, J.W., Griggs, T.C., and Wiedmeier, R.D. 2007. The value to herbivores of plant physical and chemical diversity in time and space. *Crop Science* 47:382–398.
- 52 Coppedge, B.R., Engle, D.M., Toepfer, C.S., and Shaw, J.H. 1998. Effects of seasonal fire, bison grazing and climatic variation on tallgrass prairie vegetation. *Plant Ecology* 139(2):235–246.
- 53 Helzer, C.J. and Steuter, A.A. 2005. Preliminary effects of patch-burn grazing on a high-diversity prairie restoration. *Ecological Restoration* 23(3):167–171.
- 54 Cummings, D.C., Engle, D.M., and Fuhlendorf, S.D. 2007. Is altering grazing selectivity of invasive forage species with patch burning more effective than herbicide treatments? *Rangeland Ecology and Management* 60:253–260.
- 55 Price, C.A. and Weltzin, J.F. 2003. Managing non-native plant populations through intensive community restoration in Cades Cove, Great Smoky Mountains National Park, USA. *Restoration Ecology* 11(3):351–358.
- 56 Palmer, B. 1999. Greedy grass ravaging the range experts say tenacious plant threatening areas. *The Oklahoman*. Available at Web site <http://newsok.com/greedy-grass-ravaging-the-range-experts-say-tenacious-plant-threatening-areas/article/2671769> (verified 5 December 2013).
- 57 Cook, V.K. and Hickman, K.R. 2012. Integrating the fire-grazing interaction with herbicide treatments: A novel approach to controlling *Lespedeza cuneata* in the tallgrass prairie. In Proceedings of 97th Ecological Society of America Annual Meeting. 8 August 2012. Portland, OR.
- 58 Allred, B.W., Fuhlendorf, S.D., Monaco, T.A., and Will, R.E. 2010. Morphological and physiological traits in the success of the invasive plant *Lespedeza cuneata*. *Biological Invasions* 12(4):739–749.
- 59 McDaniel, K.C., Hart, C.R., and Carroll, D.B. 1997. Broom snakeweed control with fire on New Mexico blue grama rangeland. *Journal of Range Management* 50:652–659.
- 60 Augustine, D.J. and Milchunas, D.G. 2009. Vegetation responses to prescribed burning of grazed shortgrass steppe. *Rangeland Ecology and Management* 62(1):89–97.
- 61 Strong, D.J., Ganguli, A.C., and Vermeire, L.T. 2013. Fire effects on basal area, tiller production, and mortality of the C4 bunchgrass, purple threeawn. *Fire Ecology* 9(3): 89–99.
- 62 Ansley, R.J., Boutton, T.W., Mirik, M., Castellano, M.J., and Kramp, B.A. 2010. Restoration of C4 grasses with seasonal fire in a C3/C4 grassland invaded by *Prosopis glandulosa*, a fire-resistant shrub. *Applied Vegetation Science* 13 (4):520–530.
- 63 Briggs, J.M., Hoch, G.A., and Johnson, L.C. 2002. Assessing the rate, mechanisms, and consequences of the conversion of tallgrass prairie to *Juniperus virginiana* forest. *Ecosystems* 5:578–586.
- 64 Engle, D.M., Stritzke, J.F., and Claypool, P.L. 1987. Herbage standing crop around eastern redcedar trees. *Rangeland Ecology and Management* 40:237–239.
- 65 Limb, R.F., Engle, D.M., Alford, A.L., and Hellgren, E.C. 2010. Tallgrass prairie plant community dynamics along a canopy cover gradient of eastern redcedar (*Juniperus virginiana* L.). *Rangeland Ecology and Management* 63: 638–644.
- 66 Ansley, R.J. and Rasmussen, G.A. 2005. Managing native invasive juniper species using fire. *Weed Technology* 19(3):517–522.
- 67 Teague, R., Borchardt, R., Ansley, J., Pinchak, B., Cox, J., Foy, J.K., and McGrann, J. 1997. Sustainable management strategies for mesquite rangeland: The Waggoner Kite Project. *Rangelands* 19(5):4–8.
- 68 Bidwell, T.G., Weir, J.R., and Engle, D.M. 2002. Eastern Redcedar Control and Management: Best Management Practices to Restore Oklahoma's Ecosystems F-2876. Division of Agricultural Sciences and Natural Resources, Oklahoma State University Cooperative Extension Service. Stillwater, OK.

- 69 Kerby, J.D., Fuhlendorf, S.D., and Engle, D.M. 2007. Landscape heterogeneity and fire behavior: Scale-dependent feedback between fire and grazing processes. *Landscape Ecology* 22:507–516.
- 70 Ansley, R.J., Pinchak, W.E., Teague, W.R., Kramp, B.A., Jones, D.L., and Barnett, K. 2010. Integrating grazing and prescribed fire restoration strategies in a mesquite savannah: II. Fire behavior and mesquite landscape cover responses. *Rangeland Ecology and Management* 63(3): 286–297.
- 71 Heisler, J.L., Briggs, J.M., Knapp, A.K., Blair, J.M., and Seery, A. 2004. Direct and indirect effects of fire on shrub density and aboveground productivity in a mesic grassland. *Ecology* 85:2245–2257.
- 72 Teague, W.R., Duke, S.E., Waggoner, J.A., Dowhower, S.L., and Gerrard, S.A. 2008. Rangeland vegetation and soil response to summer patch fires under continuous grazing. *Arid Land Research and Management* 22(3):228–241.
- 73 LeBauer, D. and Treseder, K. 2008. Nitrogen limitation of net primary productivity in terrestrial ecosystems is globally distributed. *Ecology* 89:371–379.
- 74 Belasco, I.J. 1954. Comparison of urea and protein meals as nitrogen sources for rumen micro-organisms: Urea utilization and cellulose digestion. *Journal of Animal Science* 13:739–747.
- 75 Anderson, R.H., Fuhlendorf, S.D., and Engle, D.M. 2006. Soil nitrogen availability in tallgrass prairie under the fire-grazing interaction. *Rangeland Ecology and Management* 59:625–631.
- 76 Knapp, A.K. and Seastedt, T.R. 1986. Detritus accumulation limits productivity of tallgrass prairie. *BioScience* 36(10):662–668.
- 77 Blair, J.M. 1997. Fire, N availability, and plant response in grasslands: A test of the transient maxima hypothesis. *Ecology* 78(8):2359–2368.
- 78 Holechek, J.L., Pieper, R.D., and Herbel, C.H. 1995. *Range Management: Principles and Practices*. Prentice-Hall, Upper Saddle River, NJ.
- 79 Bailey, D.W. 2004. Management strategies for optimal grazing distribution and use of arid rangelands. *Journal of Animal Science* 82(13):147–153.
- 80 Knight, K.B., Toombs, T.P., and Derner, J.D. 2011. Cross-fencing on private US rangelands: Financial costs and producer risks. *Rangelands* 33:41–44.
- 81 Clark, P.E., Jaechoul, L., Kyungduk, K., Nielson, R.M., Johnson, D.E., Ganskopp, D.C., Chigbrow, J., Pierson, F.B., and Hardegree, S.P. 2014. Prescribed fire effects on resource selection by cattle in mesic sagebrush steppe. Part 1: Spring grazing. *Journal of Arid Environments* 100–101: 78–88.
- 82 Clark, P.E., Jaechoul, L., Kyungduk, K., Nielson, R.M., Johnson, D.E., Ganskopp, D.C., Pierson, F.B., and Hardegree, S.P. 2015. Prescribed fire effects on resource selection by cattle in mesic sagebrush steppe. Part 2: Mid-summer grazing. *Journal of Arid Environments*, in press. doi: 10.1016/j.jaridenv.2015.03.005.
- 83 Davit, C. and Alleger, M. 2008. Patch-burn grazing benefits grassland birds and cattle producers. *Missouri Prairie Journal* 29(1):12–19.
- 84 McDonnell, M.J. 1986. Old field vegetation height and the dispersal pattern of bird-disseminated woody plants. *Bulletin of the Torrey Botanical Club* 113:6–11.
- 85 Coppedge, B.R., Engle, D.M., Fuhlendorf, S.D., Masters, R.E., and Gregory, M.S. 2001. Landscape cover type and pattern dynamics in fragmented southern Great Plains grasslands, USA. *Landscape Ecology* 16(8):677–690.
- 86 Stevens, B.S., Reese, K.P., Connelly, J.W., and Musil, D.D. 2012. Greater sage-grouse and fences: Does marking reduce collisions? *Wildlife Society Bulletin* 36(2):297–303.
- 87 Scott, M.D. 1992. Buck-and-pole fence crossings by 4 ungulate species. *Wildlife Society Bulletin* 20(2):204–210.
- 88 Johnson, L.C. and Matchett, J.R. 2001. Fire and grazing regulate belowground processes in tallgrass prairie. *Ecology* 82(12):3377–3389.
- 89 Wiens, J.A. 1974. Habitat heterogeneity and avian community structure in North American grasslands. *American Midland Naturalist* 91(1):195–213.
- 90 Winter, S.L., Fuhlendorf, S.D., Goad, C.L., Davis, C.A., Hickman, K.R., and Leslie, D.M., Jr. 2012. Restoration of the fire-grazing interaction in *Artemisia filifolia* shrubland. *Journal of Applied Ecology* 49(1):242–250.
- 91 Leis, S.A., Morrison, L.W., and Debacker, M.D. 2013. Spatiotemporal variation in vegetation structure resulting from pyric-herbivory. *Prairie Naturalist* 45:13–20.
- 92 McGranahan, D.A., Raicovich, G.M., Wilson, W.N., and Smith, C.K. 2013. Preliminary evidence that patch burn-grazing creates spatially heterogeneous habitat structure in old-field grassland. *Southeastern Naturalist* 12(3): 655–660.
- 93 McGranahan, D.A., Engle, D.M., Fuhlendorf, S.D., Winter, S.J., Miller, J.R., and Debinski, D.M. 2012. Spatial heterogeneity across five rangelands managed with pyric-herbivory. *Journal of Applied Ecology* 49(4):903–910.
- 94 Collins, S.L., and Smith, M.D. 2006. Scale-dependent interaction of fire and grazing on community heterogeneity in tallgrass prairie. *Ecology* 87(8):2058–2067.
- 95 Augustine, D.J. and Derner, J.D. 2012. Disturbance regimes and mountain plover habitat in shortgrass steppe: Large herbivore grazing does not substitute for prairie dog grazing or fire. *Journal of Wildlife Management* 76(4): 721–728.
- 96 Drewa, P.B. and Havstad, K.M. 2001. Effects of fire, grazing and the presence of shrubs on Chihuahuan desert grasslands. *Journal of Arid Environments* 48(4):429–443.
- 97 Milchunas, D.G., Sala, O.E., and Lauenroth, W.K. 1988. A generalized model of the effects of grazing by large herbivores on grassland community structure. *American Naturalist* 132(1):87–106.
- 98 Hiatt, E.L.I. 2014. Patch-burning in mixed grass prairie: Animal use of riparian areas and effects on plant species richness. Oklahoma State University Thesis, Stillwater, OK.
- 99 Allred, B.W., Fuhlendorf, S.D., Hovick, T.J., Elmore, R.D., Engle, D.M., and Joern, A. 2013. Conservation implications of native and introduced ungulates in a changing climate. *Global Change Biology* 19:1875–1883.
- 100 Vermeire, L.T., Wester, D.B., Mitchell, R.B., and Fuhlendorf, S.D. 2005. Fire and grazing effects on wind erosion, soil water content, and soil temperature. *Journal of Environmental Quality* 34(5):1559–1565.
- 101 Ozaslan, P.A., Parlak, M., Blanco-Canqui, H., Schacht, W.H., Guretzky, J.A., and Mamo, M. 2015. Patch burning: Implications on water erosion and soil properties. *Journal of Environmental Quality* 44(3):903–909.

- 102 Engle, D.M., Fuhlendorf, S.D., Roper, A., and Leslie, D.M., Jr.** 2008. Invertebrate community response to a shifting mosaic of habitat. *Rangeland Ecology and Management* 61(1):55–62.
- 103 Doxon, E.D., Winter, S.L., Davis, C.A., and Fuhlendorf, S. D.** 2011. Aboveground macroinvertebrate diversity and abundance in sand sagebrush prairie managed with the use of pyric herbivory. *Rangeland Ecology and Management* 64:394–403.
- 104 Baum, K.A. and Sharber, W.V.** 2012. Fire creates host plant patches for monarch butterflies. *Biology Letters* 8 (6):968–971.
- 105 Moranz, R.A.** 2010. The effects of ecological management on tallgrass prairie butterflies and their nectar sources. Oklahoma State University Dissertation, Stillwater, OK.
- 106 Moranz, R.A., Debinski, D.M., McGranahan, D.A., Engle, D.M., and Miller, J.R.** 2012. Untangling the effects of fire, grazing, and land-use legacies on grassland butterfly communities. *Biodiversity and Conservation* 21(11):2719–2746.
- 107 Debinski, D.M., Vogel, J.A., Koford, R.R., and Miller, J.R.** 2007. Butterfly responses to prairie restoration through fire and grazing. *Biological Conservation* 140:78–90.
- 108 Vogel, J.A., Debinski, D.M., Koford, R.R., and Miller, J.R.** 2007. Butterfly responses to prairie restoration through fire and grazing. *Biological Conservation* 140(1):78–90.
- 109 Walker, T.L. and Hoback, W.W.** 2007. Effects of invasive eastern redcedar on capture rates of *Nicrophorus americanus* and other Silphidae. *Environmental Entomology* 36(2):297–307.
- 110 Powell, A.F.L.A.** 2006. Effects of prescribed burns and bison (*Bos bison*) grazing on breeding bird abundances in tallgrass prairie. *Auk* 123(1):183–197.
- 111 Pillsbury, F.C., Miller, J.R., Debinski, D.M., and Engle, D. M.** 2011. Another tool in the toolbox? Using fire and grazing to promote bird diversity in highly fragmented landscapes. *Ecosphere* 2(3):art28.
- 112 Coppedge, B.R., Fuhlendorf, S.D., Harrell, W.C., and Engle, D.M.** 2008. Avian community response to vegetation and structural features in grasslands managed with fire and grazing. *Biological Conservation* 141(5):1196–1203.
- 113 Churchwell, R.T., Davis, C.A., Fuhlendorf, S.D., and Engle, D.M.** 2008. Effects of patch-burn management on Dickcissel nest success in a tallgrass prairie. *Journal of Wildlife Management* 72(7):1596–1604.
- 114 Hovick, T.J., Miller, J.R., Dinsmore, S.J., Engle, D.M., Debinski, D.M., and Fuhlendorf, S.D.** 2012. Effects of fire and grazing on grasshopper sparrow nest survival. *Journal of Wildlife Management* 76(1):19–27.
- 115 Hovick, T.J., Elmore, R.D., Fuhlendorf, S.D., Engle, D.M., and Hamilton, R.G.** 2014. Spatial heterogeneity increases diversity and stability in grassland bird communities. *Ecological Applications* 25:662–672.
- 116 Hovick, T.J., Elmore, R.D., and Fuhlendorf, S.D.** 2014. Structural heterogeneity increases diversity of non-breeding grassland birds. *Ecosphere* 5(5):art62.
- 117 Hernández, F. and Guthery, F.S.** 2012. Beef, Brush, and Bobwhites: Quail Management in Cattle Country. Texas A&M University Press, College Station, TX.
- 118 Bidwell, T.G., Masters, R.E., Sams, M., and Tully, S.** 2004. Bobwhite Quail Habitat Evaluation and Management Guide E-904. Oklahoma State University, Division of Agricultural Science and Natural Resources, Oklahoma State University Cooperative Extension Service, Stillwater, OK.
- 119 Robbins, M.B., Peterson, A.T., and Ortega-Huerta, M.A.** 2002. Major negative impacts of early intensive cattle stocking on tallgrass prairies: The case of the greater prairie-chicken (*Tympanuchus cupido*). *North American Birds* 56:239–244.
- 120 McNew, L.B., Preby, T.J., and Sandercock, B.K.** 2012. Effects of rangeland management on the site occupancy dynamics of prairie-chickens in a protected prairie preserve. *Journal of Wildlife Management* 76:38–47.
- 121 Griebel, R., Winter, S.L., and Steuter, A.** 1998. Grassland birds and habitat structure in sandhills prairie managed using cattle or bison plus fire. *Great Plains Research* 8: 255–268.
- 122 Fuhlendorf, S.D., Townsend, D.E., Elmore, R.D., and Engle, D.M.** 2010. Pyric-herbivory to promote rangeland heterogeneity: Evidence from small mammal communities. *Rangeland Ecology and Management* 63(6):670–678.
- 123 Augustine, D.J., Cully, J.F., Jr, and Johnson, T.L.** 2007. Influence of fire on black-tailed prairie dog colony expansion in shortgrass steppe. *Rangeland Ecology and Management* 60(5):538–542.
- 124 Breland, A.** 2010. Black-tailed prairie dog and large ungulate response to fire on mixed-grass prairie. Oklahoma State University Thesis, Stillwater, OK.
- 125 Meek, M.G., Cooper, S.M., Owens, M.K., Cooper, R.M., and Wappel, A.L.** 2008. White-tailed deer distribution in response to patch burning on rangeland. *Journal of Arid Environments* 72(11):2026–2033.
- 126 Dyke, F. and Darragh, J.A.** 2007. Response of elk to changes in plant production and nutrition following prescribed burning. *Journal of Wildlife Management* 71(1):23–29.
- 127 Hobbs, N.T. and Spowart, R.A.** 1984. Effects of prescribed fire on nutrition of mountain sheep and mule deer during winter and spring. *Journal of Wildlife Management* 48(2): 551–560.
- 128 Pearson, S.M., Turner, M.G., Wallace, L.L., and Romme, W.H.** 1995. Winter habitat use by large ungulates following fire in northern Yellowstone National Park. *Ecological Applications* 5(3):744–755.
- 129 Bleich, V.C., Johnson, H.E., Holl, S.A., Konde, L., Torres, S.G., and Krausman, P.R.** 2008. Fire history in a chaparral ecosystem: Implications for conservation of a native ungulate. *Rangeland Ecology and Management* 61(6):571–579.
- 130 Holl, S.A., Bleich, V.C., Callenberger, B.W., and Bahro, B.** 2012. Simulated effects of two fire regimes on bighorn sheep: The San Gabriel Mountains, California, USA. *Fire Ecology* 8(3):88–103.
- 131 Augustine, D.J. and Derner, J.D.** 2014. Controls over the strength and timing of fire–grazer interactions in a semi-arid rangeland. *Journal of Applied Ecology* 51:242–250.
- 132 Hovick, T.J., Elmore, R.D., Fuhlendorf, S.D., and Dahlgren, D.K.** 2015. Weather constrains the influence of fire and grazing on nesting greater prairie-chickens. *Rangeland Ecology and Management* 68(2):186–193.
- 133 Hudak, A.T., Fairbanks, D.H., and Brockett, B.H.** 2004. Trends in fire patterns in a southern African savanna under alternative land use practices. *Agriculture, Ecosystems and Environment* 101:307–325.

- 134 **Yassir, I., Van der Kamp, J., and Buurman, P.** 2010. Secondary succession after fire in Imperata grasslands of East Kalimantan, Indonesia. *Agriculture, Ecosystems and Environment* 137:172–182.
- 135 **Nayak, R.R., Vaidyanathan, S., and Krishnaswamy, J.** 2014. Fire and grazing modify grass community response to environmental determinants in savannas: Implications for sustainable use. *Agriculture, Ecosystems and Environment* 185:197–207.
- 136 **Beck, J.L., Klein, J.G., Wright, J., and Wolfley, K.P.** 2011. Potential and pitfalls of prescribed burning big sagebrush habitat to enhance nesting and early brood-rearing habitats for greater sage-grouse. *Natural Resources and Environmental Issues* 16(1):5.
- 137 **Shindler, B., Gordon, R., Brunson, M.W., and Olsen, C.** 2011. Public perceptions of sagebrush ecosystem management in the Great Basin. *Rangeland Ecology and Management* 64(4):335–343.
- 138 **Franzluebbers, A.J., Stuedemann, J.A., and Seman, D.H.** 2013. Stocker performance and production in mixed tall fescue–bermudagrass pastures of the Southern Piedmont USA. *Renewable Agriculture and Food Systems* 28(2):160–172.
- 139 **Wiltshire, K., Delate, K., Wiedenhoft, M., and Flora, J.** 2011. Incorporating native plants into multifunctional prairie pastures for organic cow–calf operations. *Renewable Agriculture and Food Systems* 26(2):114–126.
- 140 **Duru, M., Jouany, C., Le Roux, X., Navas, M.L., and Cruz, P.** 2013. From a conceptual framework to an operational approach for managing grassland functional diversity to obtain targeted ecosystem services: Case studies from French mountains. *Renewable Agriculture and Food Systems* 29(3):1–16.
- 141 **Asbjornsen, H., Hernandez-Santana, V., Liebman, M., Bayala, J., Chen, J., Helmers, M., Ong, C.K., and Schulte, L.A.** 2014. Targeting perennial vegetation in agricultural landscapes for enhancing ecosystem services. *Renewable Agriculture and Food Systems* 29(2):101–125.
- 142 **Derner, J.D., Lauenroth, W.K., Stapp, P., and Augustine, D. J.** 2009. Livestock as ecosystem engineers for grassland bird habitat in the Western Great Plains of North America. *Rangeland Ecology and Management* 62(2):111–118.
- 143 **Scasta, J.D., Engle, D.M., Talley, J.L., Weir, J.R., Fuhlendorf, S.D., and Debinski, D.M.** 2015. Drought influences control of parasitic flies of cattle on pastures managed with patch-burn grazing. *Rangeland Ecology and Management* 68(3):290–297.
- 144 **Holling, C.S. and Meffe, G.K.** 1996. Command and control and the pathology of natural resource management. *Conservation Biology* 10(2):328–337.
- 145 **Teague, W.R., Dowhower, S.L., Baker, S.A., Ansley, R.J., Kreuter, U.P., Conover, D.M., and Waggoner, J.A.** 2010. Soil and herbaceous plant responses to summer patch burns under continuous and rotational grazing. *Agriculture, Ecosystems and Environment* 137(1):113–123.
- 146 **Briske, D.D., Derner, J.D., Brown, J.R., Fuhlendorf, S.D., Teague, W.R., Havstad, K.M., Gillen, R.L., Ash, A.J., and Willms, W.D.** 2008. Rotational grazing on rangelands: Reconciliation of perception and experimental evidence. *Rangeland Ecology and Management* 61(1):3–17.
- 147 **Fuhlendorf, S.D., Engle, D.M., Elmore, R.D., Limb, R.F., and Bidwell, T.G.** 2012. Conservation of pattern and process: Developing an alternative paradigm of rangeland management. *Rangeland Ecology and Management* 65(6):579–589.